

Scotland's Rural College

## Could consumption of insects, cultured meat or imitation meat reduce global agricultural land use?

Alexander, P; Brown, C; Arneth, A; Dias, C; Finnigan, J; Moran, D; Rounsevell, MDA

*Published in:*  
Global Food Security

*DOI:*  
[10.1016/j.gfs.2017.04.001](https://doi.org/10.1016/j.gfs.2017.04.001)

First published: 22/04/2017

*Document Version*  
Publisher's PDF, also known as Version of record

[Link to publication](#)

### *Citation for published version (APA):*

Alexander, P., Brown, C., Arneth, A., Dias, C., Finnigan, J., Moran, D., & Rounsevell, MDA. (2017). Could consumption of insects, cultured meat or imitation meat reduce global agricultural land use? *Global Food Security*, 15, 22 - 32. <https://doi.org/10.1016/j.gfs.2017.04.001>

### **General rights**

Copyright and moral rights for the publications made accessible in the public portal are retained by the authors and/or other copyright owners and it is a condition of accessing publications that users recognise and abide by the legal requirements associated with these rights.

- Users may download and print one copy of any publication from the public portal for the purpose of private study or research.
- You may not further distribute the material or use it for any profit-making activity or commercial gain
- You may freely distribute the URL identifying the publication in the public portal ?

### **Take down policy**

If you believe that this document breaches copyright please contact us providing details, and we will remove access to the work immediately and investigate your claim.

# Could consumption of insects, cultured meat or imitation meat reduce global agricultural land use?

Peter Alexander<sup>1,2\*</sup>, Calum Brown<sup>1</sup>, Almut Arneth<sup>3</sup>, Clare Dias<sup>1</sup>, John Finnigan<sup>4</sup>, Dominic Moran<sup>2,5</sup>, Mark D.A. Rounsevell<sup>1</sup>

\* Corresponding author.

E-mail: [peter.alexander@ed.ac.uk](mailto:peter.alexander@ed.ac.uk)

<sup>1</sup> School of Geosciences, University of Edinburgh, Drummond Street, Edinburgh EH8 9XP, UK.

<sup>2</sup> Land Economy and Environment Research Group, SRUC, West Mains Road, Edinburgh, EH9 3JG, UK.

<sup>3</sup> Karlsruhe Institute of Technology, Institute of Meteorology and Climate Research, Atmospheric Environmental Research (IMK-IFU), Kreuzeckbahnstr. 19, 82467 Garmisch-Partenkirchen, Germany.

<sup>4</sup> The Centre for Australian Weather and Climate Research – A partnership between CSIRO and the Bureau of Meteorology, CSIRO Marine and Atmospheric Research, Canberra, Australia.

<sup>5</sup> Environment Department, University of York, York, YO10 5NG, UK.

## Acknowledgments

The research was supported by the UK's Global Food Security Programme project Resilience of the UK food system to Global Shocks (RUGS, BB/N020707/1), and the European Union's Seventh Framework Programme LUC4C (grant no. 603542). We acknowledge the support of the Scottish Government's Rural and Environment Science and Analytical Services Division funding to SRUC. Dominic Moran acknowledges support from HEFCE Catalyst-funded N8 AgriFood Resilience Programme and University of York matched funding.

# Could consumption of insects, cultured meat or imitation meat reduce global agricultural land use?

## Highlights

- Review of alternatives to conventional animal products, e.g. cultured meat and insects
- Imitation meat and insects have the highest land use efficiency
- Land use requirements are only slightly greater for eggs and poultry meat
- Cultured meat does not seem to offer substantial benefits over poultry meat or eggs
- Mix of small changes in consumer behaviour would help to achieve a sustainable diet

# 1 **Could consumption of insects, cultured meat or imitation meat** 2 **reduce global agricultural land use?**

## 3 **Abstract**

4 Animal products, i.e. meat, milk and eggs, provide an important component in global diets, but  
5 livestock dominate agricultural land use by area and are a major source of greenhouse gases.  
6 Cultural and personal associations with animal product consumption create barriers to moderating  
7 consumption, and hence reduced environmental impacts. Here we review alternatives to  
8 conventional animal products, including cultured meat, imitation meat and insects (i.e.  
9 entomophagy), and explore the potential change in global agricultural land requirements associated  
10 with each alternative. Stylised transformative consumption scenarios where half of current  
11 conventional animal products are substituted to provide at least equal protein and calories are  
12 considered. The analysis also considers and compares the agricultural land area given shifts between  
13 conventional animal product consumption. The results suggest that imitation meat and insects have  
14 the highest land use efficiency, but the land use requirements are only slightly greater for eggs and  
15 poultry meat. The efficiency of insects and their ability to convert agricultural by-products and food  
16 waste into food, suggests further research into insect production is warranted. Cultured meat does  
17 not appear to offer substantial benefits over poultry meat or eggs, with similar conversion efficiency,  
18 but higher direct energy requirements. Comparison with the land use savings from reduced  
19 consumer waste, including over-consumption suggests greater benefits could be achieved from  
20 alternative dietary transformations considered. We conclude that although a diet with lower rates of  
21 animal product consumption is likely to create the greatest reduction in agricultural land, a mix of  
22 smaller changes in consumer behaviour, such as replacing beef with chicken, reducing food waste  
23 and potentially introducing insects more commonly into diets, would also achieve land savings and a  
24 more sustainable food system.

25

## 26 **Keywords**

27 Land use; Animal products; Livestock; Dietary change; Entomophagy; Cultured meat

## 1. Introduction

Livestock provides a quarter of all the protein (and 15% of energy) consumed in food, but also creates substantial environmental impacts (FAO, 2012; Herrero et al., 2016). The area of global pasture is more than twice that of cropland, with livestock animals additionally consuming around a third of the crops harvested as feed (FAO, 2006). Despite rises in crop yields and in the efficiency of livestock production, global agricultural land area has been expanding, increasing by 464 Mha between 1961 and 2011 (Alexander et al., 2015). Land use change in recent decades has accounted for 10-12% of total anthropogenic carbon dioxide emissions, and a third since 1850 (Houghton et al., 2012; Le Quéré et al., 2015). Livestock production also contributes to atmospheric greenhouse-gas (GHG) emissions, due to methane from enteric fermentation (presently 2.1 Gt CO<sub>2</sub> eq year<sup>-1</sup> (Gerber et al., 2013)), and nitrous oxide emissions from fertiliser use on pasture and croplands in fodder production (Smith et al., 2014). In total, livestock is responsible for 12% of global anthropogenic GHG emissions (Havlík et al., 2014). A larger global population consuming a diet richer in meat, eggs and dairy (Kearney, 2010; Keyzer et al., 2005; Popkin et al., 1999; Tilman et al., 2011) has meant that agricultural land use change in the past 50 years has been dominated by the expansion of livestock production (Alexander et al., 2015). Besides the direct GHG emissions, agriculture also has large indirect emissions (e.g. from agrochemicals production and fossil fuel used) (Smith and Gregory, 2013). The combination of land use change and other emissions increases the share of agriculture in all global anthropogenic GHG emissions to between 17% and 32% (Smith and Gregory, 2013). Therefore, changing demands on agricultural production, and in particular for animal products (i.e. meat, milk and eggs), has the potential to substantially alter GHG emissions (Bustamante et al., 2014; Havlík et al., 2014). Additionally, the sparing of agricultural land provides options for further climate change mitigation measures, including afforestation or bioenergy (Humpenöder et al., 2014).

The projected rise in global population and higher per capita rates of animal product consumption, arising from higher incomes and urbanisation, suggests that livestock production will continue to increase (Tilman et al., 2011). Changes in production practices and animal genetics that increase efficiencies may help to offset some of the potential land use and associated environmental impacts (Havlík et al., 2014; Le Cotty and Dorin, 2012). Nevertheless, demand-side measures to reduce animal product consumption may be necessary to meet climate change targets (UNFCC, 2015), while helping to achieve food security (Bajželj et al., 2014; Lamb et al., 2016; Meadu et al., 2015; Smil, 2013). High levels of meat consumption are also detrimental to human health, with links to obesity, cardiovascular diseases and cancer (Bouvard et al., 2015; Hu, 2011; NCD Risk Factor Collaboration, 2016; Popkin and Gordon-Larsen, 2004). Despite both the health and environmental benefits, changing consumer preferences towards a low meat diet is difficult because of cultural, social and personal associations with meat consumption (Graça et al., 2015; Macdiarmid et al., 2016). Although there is some evidence for increasing rates of vegetarianism and reduced meat diets in western countries (Leahy et al., 2011; Vinnari et al., 2010), the global average per capita rate of animal product consumption has continued to increase (FAOSTAT, 2015a).

Studies of the food system that include the impact of dietary change typically assume the continuation of existing consumption patterns and income and price elasticity relationships (e.g. Engström et al., 2016a, 2016b; Schmitz et al., 2014; Tilman et al., 2011), implicitly discounting the possibility of major shocks or transformative changes in diets. There has also been an increasing number of studies considering the impact of alternative assumptions regarding future diets, such as

1 lower animal product consumption, healthy diets, vegetarianism or veganism, e.g. (Bajželj et al.,  
2 2014; Erb et al., 2016; Haberl et al., 2011; Mora et al., 2016; Popp et al., 2010; Stehfest et al., 2009)

3  
4 However, technology changes or radical alteration of consumer preferences, which could be  
5 transformative for the food system, remain unexplored. New technologies raise the possibility of  
6 supplying high quality food from novel sources, e.g. cultured meat, also known as *in vitro* meat  
7 (Thornton, 2010). Also, behaviour, preferences and social norms change over time, such that food  
8 previously considered unacceptable or undesirable (e.g. insects, in western countries) could become  
9 a more common part of future diets (van Huis, 2013). There are historical precedents for foods  
10 becoming acceptable after long periods of rejection; for example, tomatoes in Britain were widely  
11 viewed with suspicion and dismissed for over 200 years (Bir, 2014; K. A. Smith, 2013). Similarly,  
12 lobster in America was initially a poverty food eaten by slaves and prisoners, and used as fertiliser  
13 and fish bait, due to their abundance (Dembosky, 2006). It wasn't until the late nineteenth century  
14 that lobster developed a status as a luxury food, supported by the expansion of the US railway  
15 network giving access to new markets (Townsend, 2012). But while alternative food sources may  
16 become technologically feasible or publically acceptable in the future, their potential contributions  
17 to sustainability remains unclear.

18  
19 This study addresses this research gap by reviewing and comparing the potentially transformative  
20 alternatives to conventional animal products, including cultured meat, imitation meat and insects,  
21 and consider the implications for global agricultural land use requirements given widespread  
22 adoption. The approach is explorative, rather than predictive, and assumes half of existing animal  
23 products are substituted by each alternative food, to provide at least equal energy and protein. The  
24 objective is to compare the alternatives on an equal basis and to assess their potential to reduce  
25 agricultural land requirements, and contribute to food system sustainability. To allow comparison  
26 with more typical dietary change, several other scenarios were also included using the same  
27 methodology. These scenarios include shifts in conventional animal product consumption, changes  
28 to high and low animal product diets (based on average consumption in India and the USA), and  
29 reductions in consumer waste. The focus is on animal products due to their dominance in the food  
30 system for land use and environmental impacts (Herrero et al., 2016), and because of their relative  
31 inefficiency in converting inputs into human-edible food (FAO, 2006; Mottet et al., 2017). The  
32 premise is that due to the cultural and personal associations with animal product consumption  
33 (Graça et al., 2015; Macdiarmid et al., 2016), consumers with higher incomes continue to eat large  
34 quantities of animal products and consumers currently eating at lower rates will increase their  
35 consumption as incomes increase. This assumption combined with population growth, also  
36 underlies the projections of substantial increases (from 76 to 133%) in global animal product demand  
37 (Alexandratos and Bruinsma, 2012; Bodirsky et al., 2015). Therefore, alternatives that mimic aspects  
38 of these products in a manner that is acceptable to consumers need to be explored for  
39 environmental sustainability.

## 40 41 **2. Alternatives to current animal products**

42 There are several alternatives to existing animal products as food protein and energy sources:

1 (a) *Insects*

2 Edible insects have the potential to become a major source of human nutrition, and can be produced  
3 more efficiently than conventional livestock, i.e. in terms of converting biomass into protein or calories  
4 (Tabassum-Abbasi et al., 2016; van Huis, 2013). They are high in fat, protein and micronutrients  
5 (Persijn and Charrondiere, 2014; Rumpold and Schlüter, 2013), and can be produced with lower levels  
6 of GHG emissions and water consumption (van Huis, 2013). The efficiency of insects to convert feed  
7 into edible food is in part due to the higher fraction of insect consumed (up to 100%), compared to  
8 conventional meat (e.g. 40% of live animal weight is consumed with cattle). Insects are poikilothermic,  
9 so they do not use their metabolism to heat or cool themselves, reducing energy usage. They tend to  
10 have higher fecundity than conventional livestock, potentially producing thousands of offspring  
11 (Premalatha et al., 2011). Efficiency is also increased by rapid growth rates and the ability of insects  
12 to reach maturity in days rather than months or years.

13

14 Isotope analysis of bones indicates that insectivorous diets are entrenched in human evolution (De-  
15 Magistris et al., 2015; Ramos-Elorduy, 2009), and a variety of species are currently consumed (>2000  
16 species (Rumpold and Schlüter, 2013)) across many regions of the world (119 countries (Rumpold and  
17 Schlüter, 2013)). But issue of limited consumer acceptability is prevalent particularly in western  
18 countries. These are also the countries with high animal product consumption rates per capita, and  
19 are therefore where a switch from animal product to insect consumption would have the greatest  
20 impact. There are already signs that consumer attitudes in developed countries such as the USA and  
21 the UK may be starting to change (Jamieson, 2015), and there may be less of a barrier to including  
22 insect-derived materials in other products, for example in powdered form (Little, 2015). However, in  
23 some jurisdictions, there are legal barriers. For example, within the European Union, regulations on  
24 novel food and the legal status of insect-based foods means that insects cannot be processed, and  
25 must be marketed whole (De-Magistris et al., 2015).

26

27 (b) *Cultured meat*

28 Cultured meat, also termed *in vitro*, 'lab-based', or synthetic meat, refers to meat produced outside of  
29 a living animal. The meat is produced by culturing animal stem cells in a medium that contains  
30 nutrients and energy sources required for the division and differentiation of the cells into muscle cells  
31 that form into tissue (Bhat et al., 2015), with commercial scale production anticipated by 2021  
32 (Verstrate, 2016). The tissue produced can be separated for further processing and packaging. The  
33 amount of nutrients and energy needed may be relatively small, as only muscle tissue develops,  
34 without the need for biological structures such as respiratory, digestive or nervous systems, bones or  
35 skin (Bhat et al., 2014). Rapid growth rates mean that tissue is maintained for a shorter time than for  
36 animal rearing, further reducing required inputs.

37

38 Cell and tissue culture are currently not efficient processes in terms of energy, water and feedstock  
39 expenditure, and have been primarily employed in scientific and medical applications (Moritz et al.,  
40 2015). The financial and sustainability advantages are also unclear as the reductions in some inputs  
41 may be offset by the extra costs of a stricter hygiene regime and other energy inputs (Bhat et al., 2014).  
42 The cell culture medium can be produced from materials of animal origin (e.g. bovine serum), but this  
43 defeats many of the sustainability benefits of cultured meat (Bhat et al., 2014). Although suitable  
44 culture medium can be produced from non-animal sources (e.g. hydrolysed cyanobacteria, sometimes  
45 known as blue-green algae (Tuomisto and de Mattos, 2011) and Maitake mushroom extract (Bhat et

1 al., 2014)), an efficient process to manufacture animal-free media is still viewed as a major challenge,  
2 and a barrier to cultured meat adoption (Mattick et al., 2015a). Consumer perceptions are also a  
3 potential barrier (Hocquette, 2016). The product needs to be of sufficiently similar taste, texture and  
4 appearance to livestock meat for wide acceptance, and this is currently difficult to achieve (Moritz et  
5 al., 2015).

### 6 7 *(c) Imitation meat*

8 Imitation meat or meat analogues attempt to mimic specific types of meat, including the aesthetic  
9 qualities (e.g. texture, flavour and appearance) and the nutrient qualities, without using meat  
10 products. Soy based products, such as tofu or tempeh, are perhaps the most widely known imitation  
11 meats (Malav et al., 2015). Tofu is soybean curd, made from coagulated soy milk, and has been  
12 prepared and consumed in Asia for centuries. It can be further prepared to approximate meat  
13 products in flavour and texture, e.g. with flavouring added to make it taste like chicken, beef, lamb,  
14 ham or sausage (Malav et al., 2015). Soy and tofu contain high levels of protein, while being low in fat  
15 (Sahirman and Ardiansyah, 2014). Beef and soy have a similar Protein Digestibility–Corrected Amino  
16 Acid Score (PDCAAS), indicating that they have similar protein values in human nutrition (Schaafsma,  
17 2000). More recent imitation meats include mycoprotein-based Quorn (Finnigan et al., 2010), and  
18 textured vegetable protein, again often made from soy.

### 19 20 *(d) Aquaculture*

21 Global aquaculture is already a major source of food, and has grown substantially over the past 50  
22 years to produce around 61.9 Mt in 2011 (FAO, 2016), which is similar to the quantity of bovine meat  
23 (FAOSTAT, 2015b). As a global per capita average, protein from fish contribute 10% (2.72  
24 g/capita/day) of that from meat, milk and eggs; 27.69 g/capita/day (FAOSTAT, 2015b), around half of  
25 which is from aquaculture. Asia dominates aquaculture production (accounting for 89 per cent by  
26 mass), with 62.4% produced in China alone, due to pre-existing aquaculture practices and a relaxed  
27 regulatory framework (Bostock et al., 2010). Carnivorous fish, such as salmon, can consume up to 5  
28 times the quantity of fish (as feed) than they ultimately provide (Naylor et al., 2009). Therefore,  
29 limitations on the sustainable sourcing of feed represents a barrier to increases in farmed  
30 carnivorous fish (Diana, 2009), making substantial substitution with existing animal products less  
31 likely. This issue is less acute for herbivorous and omnivorous species, as they have much lower ‘fish-  
32 to-fish’ conversion ratios, e.g. carp currently has a ratio of 0.1, with further reductions predicted  
33 (Tacon and Metian, 2008) as fish derived feed consumption is not essential for their nutrition  
34 (Bostock et al., 2010). Freshwater aquacultural systems dominate production, accounting for around  
35 two thirds of all outputs from aquaculture. The main species are herbivorous or omnivorous, with  
36 largest production from carp, although tilapia and catfish production have increased more recently  
37 (Bostock et al., 2010).

## 38 39 **3. Comparison of land requirements**

40 To provide an assessment of the consequences of adoption of above alternative protein sources on  
41 agricultural land requirements separate scenarios for each were considered, assuming replacement  
42 of 50% of current animal products. These scenarios assume that perceptions and diets alter over  
43 time, such that current animal product (i.e. meat, milk and eggs) consumption declines and is  
44 substituted by a replacement food that provides nutritional content at least as equal in both energy



1 and protein terms. The 50% replacement assumption is largely arbitrary, but is simply used as a  
2 reference point against which to compare alternative diets. It would have been equally accurate to  
3 select an alternative value, and the relative changes between these substitution scenarios would not  
4 have been impacted, i.e. the changes would scale proportionately. Further scenarios considered  
5 conventional animal products in the same manner (i.e. 50% replacement), to provide a basis for  
6 comparison with the transformative scenarios. The scales of animal product substitution tested is  
7 not highly relevant, but rather the comparative outcomes between the substitution scenarios. The  
8 scenarios of reduced consumer waste (including both food waste and consumption in excess of  
9 nutritional requirements) and global adoption of the current average per capita diets in India and the  
10 United States of America were also constructed. These scenarios are not chosen to be equally  
11 probable or desirable, but rather to provide a broad comparison between the impacts of potential  
12 transformations in consumer behaviour.

13

### 14 **3.1. Human appropriation of land for food**

15 Results are expressed using the Human Appropriation of Land for Food (HALF) index (Alexander et al.,  
16 2016), giving the percentage of global land surface required to supply the world's population with a  
17 particular diet, under current production efficiencies. The baseline 2011 HALF index was calculated  
18 from FAO country-level panel data for crop areas, production quantities, commodity uses and nutrient  
19 values (FAOSTAT, 2015a, 2015c, 2015d, 2015e, 2015f, 2015g). Following the approach of Alexander *et*  
20 *al.* (2016), 90 commodities (50 primary crops that are directly grown, 32 processed commodities  
21 derived from them, and 8 livestock products (2016)), representing 99.4% of global food consumption  
22 by calorific value, were considered.

23

24 The areas associated with primary crops production were determined using yields adjusted to include  
25 losses in storage and transport (overall around 5%), calculated by *pro rata* allocation of these losses to  
26 subsequent uses. These yields were multiplied by the quantity of each commodity used as food for  
27 human consumption, processing and animal feed (FAOSTAT, 2015a, 2015d) to obtain an associated  
28 production area. The areas for the processed primary crops were mapped to the commodities  
29 produced, and allocated by economic value (e.g. soybeans processed into soybean oil and meal)  
30 (Alexander et al., 2016). The feed use was divided between animal products using estimated feed  
31 requirements. Monogastric livestock (i.e. poultry and pigs) nutrition was assumed to be met solely  
32 from feed, while feed and grazed pasture is used for ruminant species (e.g. cattle and sheep). Feed  
33 requirements were calculated using feed conversion ratios (FCRs), which express the efficiency of  
34 converting biomass inputs into animal products (Little, 2014; Macleod et al., 2013; Opio et al., 2013;  
35 Smil, 2013). The feed requirements for monogastrics were assigned first, and remaining feed and the  
36 total pasture area were then allocated *pro rata* by feed requirements to the ruminant products.

37

38 This approach provides the yields for primary crops, processed commodities and livestock products,  
39 using 2011 global average production efficiencies. These were used to estimate the cropland and  
40 pasture areas needed for diets containing these commodities, with the resulting areas expressed  
41 through the HALF index, i.e. as the percentage of total land area required for food production. The  
42 HALF index does not provide a land use footprint for particular countries or regions, but addresses  
43 questions such as "how much land would be used if the global population adopted diet X". The  
44 approach provides a comparative metric of the land requirements of different diets, and a way to  
45 consider the impacts from changes in dietary patterns. The inclusion of local production systems

1 within a land footprint would tend to obscures the understanding of the role of diet in the global food  
2 system.

3

### 4 ***3.2. Alternative animal product scenarios***

5 The alternative animal product scenarios assume that 50% of current animal products, evenly  
6 distributed across existing sources, are replaced by one commodity, while being constrained to  
7 maintaining at least equal quantities of energy and protein within the diet. Nutrient contents and  
8 FCRs were estimated for the substitute commodities (Table 1, with assumptions below). The protein  
9 and energy contents were used to calculate the mass of the commodity required to replace the  
10 conventional foods removed. FCRs were applied to evaluate the feed requirements to produce the  
11 substitute product. The feed was assumed to be provided from the current mix and yields of animal  
12 feeds, except for imitation meat, which was calculated using soybean production. The net changes in  
13 cropland and pasture areas were then calculated assuming the conventional livestock area reduces  
14 by 50% (assuming constant production practices) plus the requirements from the replacement  
15 commodity.

16

1 Table 1. Feed conversion efficiencies, in dry matter (DM) weight of feed required per unit edible  
 2 weight (EW), for alternatives to convention animal products considered. For conventional livestock  
 3 feed conversion efficiencies data used and sources are given in Table 1, Alexander et al. (2016).

Commodity	Percentage edible (% EW of LW)	Feed conversion by mass (kg DM feed/kg EW) [uncertainty range]	Energy content (MJ/kg EW)	Protein content (g / kg EW)	Energy feed conversion efficiency <sup>a</sup> (%)	Protein feed conversion efficiency <sup>a</sup> (%)	Direct energy for housing and processing (MJ / kg EW)	Data sources
Mealworm: larvae ( <i>Tenebrio molitor</i> )	100	1.8 <sup>b</sup> [1.6-2.1]	8.9	179	33	50	7.3	(Oonincx and de Boer, 2012; Persijn and Charrondiere, 2014; Spang, 2013)
Crickets: adults ( <i>Acheta domesticus</i> )	80	2.1 [1.9-2.4]	5.9	205	19	49	No data	(Finke, 2002; van Huis, 2013)
Cultured meat	100	4 [2-8]	8.3	190	17	24	18-25 <sup>c</sup>	(Tuomisto and de Mattos, 2011)
Imitation meat (based on soybean curd) <sup>d</sup>	-	0.29 [0.27-0.35]	3.2	81	47	72	11.4	(Sahirman and Ardiansyah, 2014; USDA, 2015; Wang and Cavins, 1989)
Tilapia	37	4.6 [3.7-5.5]	4.0	201	5.8	21.8	5.4	(Pelletier and Tyedmers, 2010; USDA, 2015)
Chinese Carp	37	4.9 [3.9-5.9]	5.3	178	7.3	18.3	5.4 <sup>e</sup>	(Bauer and Schlott, 2009; Tacon and Metian, 2008; USDA, 2015)

**Notes:**

- Energy and protein conversion efficiency based on feed content of 15 MJ/kg DM and 200 g/kg protein.
- Mealworm feed efficiency adjusted from Spang (2013), assuming 62% moisture content (Persijn and Charrondiere (2014).
- Excluding production of biomass feedstock.
- Feed columns relates to inputs of soy to tofu production process.
- Based on Tilapia production.

4

5 (a) *Insect consumption*

6 Mealworm larvae and adult crickets were selected to assess the impact of insect consumption, based  
 7 on the availability of data for these species (Table 1). Protein from conventional livestock and insects  
 8 were considered substitutable on an equal mass basis, as all essential amino acids for humans are  
 9 available from insects, although profiles differ between species (Persijn and Charrondiere, 2014; van  
 10 Huis, 2013). Insects are also high in a variety of micronutrients such as the minerals copper, iron,  
 11 magnesium, manganese, phosphorous, selenium, and zinc and the vitamins riboflavin, pantothenic  
 12 acid, biotin, and in some cases folic acid (Persijn and Charrondiere, 2014; Rumpold and Schlüter, 2013).  
 13 However, the analysis is limited to considering equivalence of protein and energy only. Although  
 14 insects can be produced from organic wastes, given the high levels of production required under this  
 15 scenario, it is assumed that production is from purpose-grown feed, rather than waste sources.

16

1 *(b) Cultured (in vitro) meat*

2 Process efficiency values from Tuomisto & de Mattos (2011) were used as FCR, but assuming that the  
3 raw materials for the production of the culture medium is from conventional livestock feeds (Table 1).  
4 Tuomisto & de Mattos (2011) suggest 99% less land is required to produce cultured meat rather than  
5 livestock meat, but this assumes production of biomass for the culture medium using an algae-based  
6 system. This increases direct energy requirements while reducing land requirements, but depends  
7 upon a conflation of two novel technologies; production of algae biomass and cell culturing of meat.  
8 Producing feed from algae is likely to reduce the land required for conventional livestock production,  
9 while increasing other inputs, and therefore we consider only the cultured meat aspect. Production  
10 of the nutrient 'broth' in which the cells are cultured (Mattick et al., 2015a; Verbeke et al., 2015) is  
11 possible from different inputs. However, as commercial-scale processes for cultured meat are not yet  
12 available (Mattick et al., 2015a), the assessment of which feedstock would be selected to produce the  
13 culture media in the required quantities, and the associated efficiency are both uncertain. To  
14 represent this uncertainty the conversion efficiency range tested is large (Table 1).

15

16 *(c) Imitation meat*

17 The calculation was based on the use of soybean curd, i.e. tofu, for imitation meat. Manufacturing  
18 soybean curd from soybeans creates some losses in protein and energy content (Wang and Cavins,  
19 1989), for example during the washing, grinding, boiling and pressing involved (Sahirman and  
20 Ardiansyah, 2014), and also requires direct input of energy to these operations (Table 1). The  
21 production of the soybean curd was considered analogously to livestock production, with soy being  
22 used to produce soybean curd, rather than livestock inputs producing animal products. The losses in  
23 preparation of imitation meat from the soybean curd are expected to be low, and given the relatively  
24 simple processes, such as extrusion (Malav et al., 2015), have substantially lower direct energy inputs  
25 in comparison to cultured meat.

26

27 *(d) Aquaculture*

28 Production of Chinese carp and tilapia were taken as examples in the analysis, due to their high  
29 contribution to current aquaculture and, compared to carnivorous fish (e.g. salmon), their low  
30 requirements for fishmeal or fish oil as feeds, and more advantageous FCR. The feed conversion  
31 ratios to live weight for tilapia and carp are 1.7 and 1.8 respectively (Tacon and Metian, 2008), but  
32 given that only 37% of the fish by weight is fillet (Bauer and Schlott, 2009; Pelletier and Tyedmers,  
33 2010), this leads to a FCR to edible weight of 4.6 to 4.9 (Table 1). Although some fishmeal and fish oil  
34 are currently used as feed for these species, these are not essential for nutrition in herbivorous and  
35 omnivorous species (e.g. carp and tilapia) (Bostock et al., 2010). Therefore, the assumption is that all  
36 feed is provided from land-based production (e.g. soybeans and cereals). Any contribution from  
37 fishmeal and fish oil, that could be provided sustainably from fish processing by-products is  
38 neglected (Bostock et al., 2010; Tacon and Metian, 2008). The 50% replacement scenario would  
39 imply an approximately 10-fold increase in protein terms.

40

41 *(e) Conventional livestock consumption changes*

42 Each of the conventional animal products was also considered as replacements for 50% of the  
43 current mix. Thus, more than half of calories or protein were assumed to be provided by the  
44 commodity being considered in each of these scenarios. For example, poultry meat currently  
45 provides 24% of all animal proteins, which reduces to 12% under all the other protein meat

1 substitution scenarios except the poultry meat scenario. Under this scenario 62% of animal product  
2 consumption is from poultry, i.e. the 12% of unchanged poultry consumption plus the 50%  
3 substituted for the current animal product mix. The feed and pasture area requirements were  
4 calculated using the results derived from the FAO data (FAOSTAT, 2015a, 2015c, 2015d, 2015e,  
5 2015f, 2015g), as described above.  
6

### 7 **3.3. Waste and other dietary change scenarios**

8 Scenarios for food waste reduction and for global adoption of the average diets in India and the USA  
9 were included from previously calculated results (Alexander et al., 2017, 2016).  
10

#### 11 *(a) Waste reduction*

12 The waste reduction scenario uses losses from Alexander *et al.* (2017). The scenario assumes that the  
13 combination of food discarded by consumers and due to over-consumption halves from the 2011 rates  
14 to 11% of energy and 26% of protein (assuming requirements of 9.8 MJ/person/day of energy and 52  
15 g/day of protein (Institute of Medicine, 2005; SACN, 2011)). The reduction in this waste was applied  
16 equally across all commodities. Losses during production, processing and distribution were not  
17 changed, as the focus here is on the impact of consumer behaviour on the food system.  
18

#### 19 *(b) High and low animal product diets*

20 To assess the impact of diets with high and low rates of animal products consumption, the average  
21 per capita consumption in the USA and India were chosen, respectively. Alexander *et al.* (2016) used  
22 the average consumption per capita for each commodity in these countries to calculate the HALF  
23 index for their diets. Additionally, the difference between the global average diet and the diets in  
24 each of the countries was decomposed into two parts (Alexander et al., 2016). The first represents a  
25 shift in the total quantity of nutrients consumed while holding the proportional contribution of each  
26 commodity constant. The second represents a shift in the profile of commodities consumed, while  
27 holding the total nutrient level constant.  
28

### 29 **3.4. Uncertainty quantification**

30 A number of the parameter values used are uncertain, with perhaps the most influential ones being  
31 the livestock feed conversion ratios and the food nutrient contents. To assess the impact of these  
32 uncertainties, these parameters were randomly sampled from assigned uncertainty ranges (i.e. a  
33 Monte Carlo uncertainty method). The range of FCR for conventional livestock was taken as -20% to  
34 +20% of the assumed value (Alexander et al., 2016, Table 1), and for the alternative commodities the  
35 ranges are given in Table 1. The ranges for protein and energy contents were -10% to + 10% for the  
36 90 agricultural commodities, carp, tilapia, soybean curd and cultured meat. However, the nutrient  
37 content of the insect species appears to be less certain, so a -30% to + 30% range was used. All of  
38 these uncertainty ranges are indicative of qualitative levels of confidence in the default values used  
39 in the absence of relevant quantitative data. Uniform distributions were used for all parameter  
40 uncertainties, sampled 500 times. The initial allocation of land use to commodities, using the  
41 methodology of (Alexander et al. (2016)), was re-run for each sampled set of FCR.  
42

### 3.5. Yields of alternatives to animal product

The energy and protein produced per unit of agricultural area were found to vary by more than 100-fold across conventional animal products and the alternatives considered (Figure 1). Soybean curd had the highest energy and protein yields ( $2.2 \text{ MJ/m}^2$  and  $57 \text{ g/m}^2$ ) and beef the lowest ( $0.02 \text{ MJ/m}^2$  and  $0.4 \text{ g/m}^2$ ). After soybean curd, the two insect species gave the next highest yields. The yields for cultured meat were similar to eggs, and also relatively close to those for poultry. The order of commodities by yield differed between protein and energy, due to the differences in nutrient contents. For example, tilapia has a higher protein, but a lower energy yield, than carp. The areas for the ruminant derived products (i.e. mutton and goat meat, milk, and beef) include both cropland to produce feed and pasture area for grazing, while the other products use only feeds from cropland.

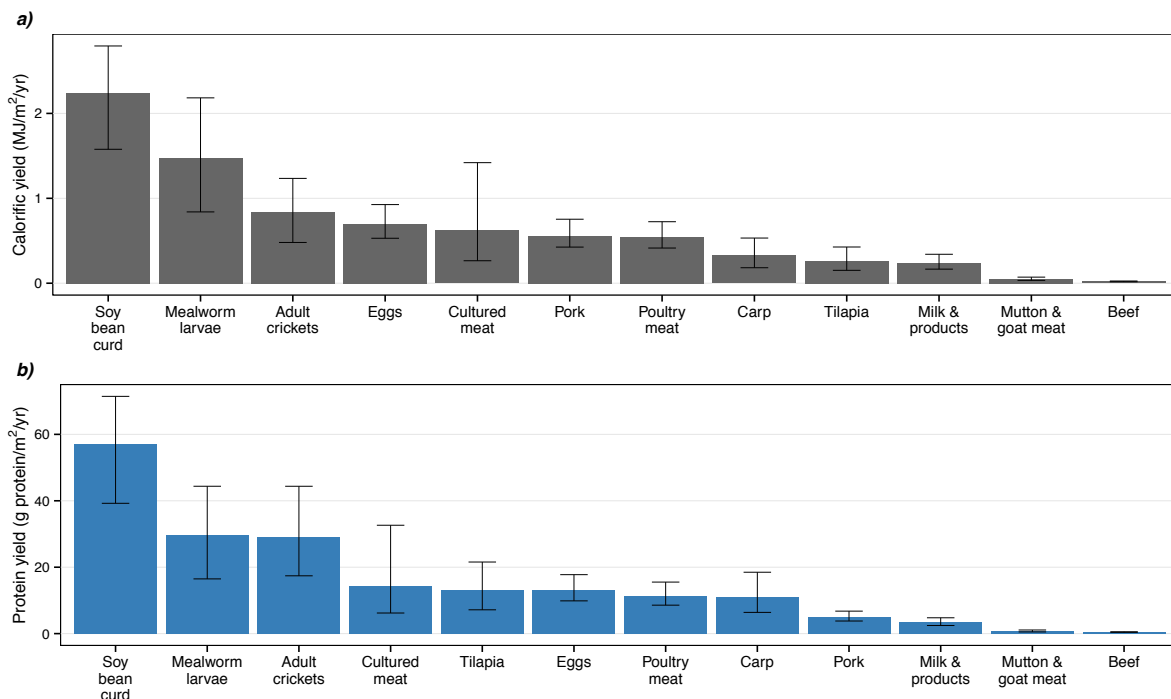


Figure 1. Energy and protein per unit area of agricultural land for conventional and alternatives to animal production. Error bars show the yield range from uncertainty in feed conversion ratios and nutrient contents.

### 3.6. Land requirements of scenarios

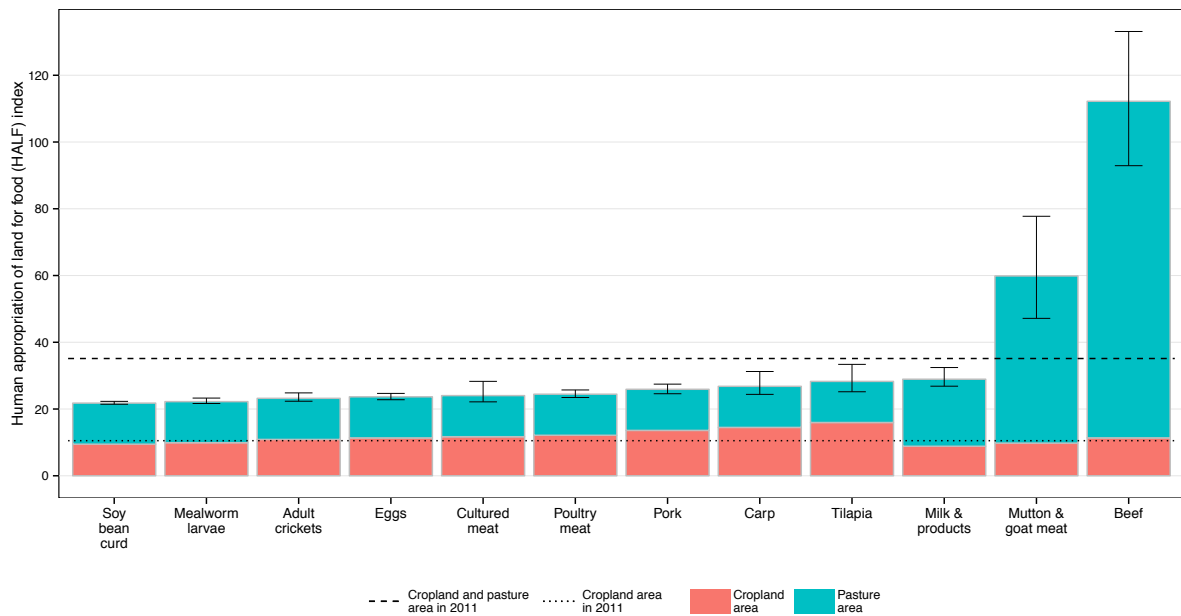
Global cropland and pasture areas vary substantially under the scenarios (Figure 2). The animal product substitute scenarios suggest that the HALF index (i.e. the percentage of land area required for food production), is 21.8 for soybean curd, and 112.2 for beef, compared to a baseline of 35.1 in 2011. There is also considerable variability in the cropland areas. The highest cropland requirement occurs in the tilapia scenario, where an additional 709 Mha of cropland is needed for feed, a 46% increase in the total cropland area. However, total agricultural land area reduces by 18% or 892 Mha, as cropland increases are more than offset by a 1601 Mha drop in pasture area. For the animal product replacement scenarios, the lowest cropland area is for milk with the cropland reducing by 217 Mha (14%) of cropland and 590 Mha (18%) of pasture, due to higher feed conversion ratios than the current mix of animal products, and because nutrients are also derived from pasture. Pasture changes dominate the results, with the cropland changes for most of the other scenarios being more

1 modest. For example, the results with the largest agricultural area change have only a 7-9% change  
 2 in cropland, with soybean curd decreasing by 137 Mha and beef increasing by 110 Mha, while the  
 3 pasture areas decrease by 1601 Mha and increase by 9916 Mha, respectively.

4  
 5 The animal production replacement scenarios all provide at least the same amount of both energy  
 6 and protein. The binding constraints were by energy for all scenarios except pork. In these scenarios  
 7 the replacement food provides an equal amount of energy, but a greater quantity of protein.

8 Conversely, for pork the binding constrained was on protein, due to the relatively low ratio of protein  
 9 to energy in pork compared to the other animal products (FAOSTAT, 2015e).

10



11

12 *Figure 2. Total cropland and pasture areas for food production under scenarios assuming 50% of*  
 13 *current nutrients from animal productions are substituted with the indicated food, to provide at least*  
 14 *equal energy and protein. The results are expressed as the percentage of global land required, or*  
 15 *HALF index, based on 2011 population and food production systems. Error bars show the HALF range*  
 16 *from uncertainty in feed conversion ratios and nutrient contents.*

17

18 The range of agricultural land areas required based on uncertainty in FCRs and food nutrients (Figure  
 19 2, error bars) are small for the animal product scenarios with low HALF indices (e.g. soybean curd and  
 20 insects). This is because the uncertainty from new food commodities, e.g. for soybean curd, is only a  
 21 small proportion of the total agricultural area, therefore a large percentage uncertainty (Figure 1)  
 22 only produces a small absolute uncertainty in land area (Figure 2). The opposite is the case for the  
 23 results with higher HALF (e.g. beef), where the areas for replacement production are large and so,  
 24 therefore, are the associated uncertainties. Figure 1 shows uncertainty for each scenario per unit of  
 25 energy or protein.

26

27 The similarity in land requirements between the commodities with low HALF indices (Figure 2)  
 28 suggests that substantial land use and associated environmental benefits could be achieved from the  
 29 adoption of any of them individually or in combination. Land requirements are always reduced by

1 further increases in efficiencies of production per unit area. For example, a doubling of efficiency  
2 between two alternative scenarios always produces a halving of land use requirements. However, as  
3 the land use requirements decrease, the differences in the absolute areas also decrease, creating  
4 diminishing returns from increasing efficiency. The selection of the most appropriate mix of the  
5 more efficient products (Figure 2) may therefore be more greatly influenced by other production  
6 externalities, e.g. biodiversity or water usage, rather than the land requirements.

7  
8 Table 2 summarises these meat substitution scenario results and also includes the results from the  
9 consumer waste and scenarios from adoption of high and low animal product diets (based on  
10 average consumption in India and the USA) (Alexander et al., 2017, 2016). As these additional  
11 scenarios involve different assumptions, i.e. they do not consider a 50% substitute of animal  
12 products, direct comparisons between these two scenario groups must be limited. However, the  
13 high and low animal product diets (based on USA and India), respectively, were found to have higher  
14 and lower land impacts than the meat alternatives, with the exception of beef (Table 2). This is  
15 because the diets include both a shift in the amounts of food consumed and, more importantly, in  
16 the types of food consumed (Alexander et al., 2016). These diets involve different rates of meat  
17 consumption, and therefore are not restricted to maintain 50% of the current animal products as in  
18 the other scenarios. The consumer waste scenario, halving foods discarded and lost due to over-  
19 consumption, was found to spare 9% of agricultural land.

20



1 *Table 2. Summary of results across all scenarios, ordered by increasing agricultural land use.*

Scenario	Description	Percentage change in required agricultural area for food	HALF index	Comments
Low animal product diet	Average diet globally becomes that of the average diet in India	-55	15.7	Influenced by lower overall consumption, and lower rates of meat in the diet. In both these aspects global diets are changing in the opposite direction of current trends, making this scenario of low plausibility.
Soybean curd	Soybean curd replaces 50% of current animal products	-35	21.7	Increase in direct energy inputs in comparison to animal products, but less substantial than for cultured meat. 50% uptake seems unlikely to be acceptable to consumers.
Insects	Mealworm larvae replaces 50% of current animal products	-34	22.2	Consumer acceptability barriers in some regions. A lower level of uptake in combination, perhaps as an ingredient, e.g. in pre-packaged foods, seems more likely.
Most efficient conventional animal products	Eggs or chicken replaces 50% of current animal products	-30 to -28	23.7 to 24.4	The direction of recent changes, with rapid growth in the consumption rates for chicken in particular, supported by intensification in production.
Cultured meat	Cultured meat replaces 50% of current animal products	-29	24.0	Technology still rather uncertain (Bhat et al., 2014), and benefits compared to other sources of nutrients currently are not well demonstrated. The high direct energy used in production also a concern.
Most efficient aquacultural product	Carp replaces 50% of current animal products	-22	26.8	Potential for environmental pollution issues with large-scale production, although this is also the case with other intensive animal production.
Milk and products	Milk and products replaces 50% of current animal products	-16	28.9	Associated with the largest reduction of cropland, while still providing material reduction in overall agricultural area.
Reduction in waste	Consumer waste, including food discard and due to over-consumption is halved	-9	32.0	Feasible, but opposite to current direction of change, particularly with respect to over-consumption. Health, as well as environmental, benefits for policies or social changes to reverse these changes.
High animal product diet	Average diet globally becomes that of the average diet in the USA	+178	97.7	Not possible given production systems currently used. Direction of recent changes for overall nutrients and rates of animal products consumption. Approaching this consumption globally would be expected to increase food price, suppress demand and intensify production practices.
Least efficient conventional animal product	Beef replaces 50% of current animal products	+204	112.2	Physically impossible with production systems currently used, and contrary to current trends of average per capita consumption falling since 1970s.

2

### 3 **4. Discussion**

#### 4 *(a) Limitations of the analysis*

5 A stylised and exploratory approach is used to better understanding and ensure comparison on a  
6 like-for-like basis of potential land use outcomes across a range of scenarios, from the more unusual  
7 and transformational (e.g. insects and cultured meat), to the more conventional (e.g. changes in  
8 proportions of livestock demand). The replacement of at least equal quantities of protein and  
9 calories has been considered, leaving the potential for reductions in micronutrients between the

1 scenarios. The results are not intended as predictive, nor are they presented to suggest equal  
2 plausibility, but rather to allow comparisons in land use requirements between the scenarios.

3  
4 Fixed global average production figures based on 2011 were used and no spatial variation in  
5 production practices are taken into account. These production practices would be expected to  
6 respond to the substantial changes considered in these scenarios, mediated by international trade in  
7 agricultural commodities. For example, increased agricultural land requirement would tend to  
8 intensify production, with higher rates of inputs used to achieve greater yields. Conversely, if less  
9 agricultural land is needed for food, this may cause a lowering of the production intensity. In both  
10 cases, such adaptation in production moderates the land use consequences, but alters the resource  
11 requirements for other inputs, e.g. fertiliser or pesticide use (Hertel et al., 2016; P. Smith, 2013).  
12 However, the results do characterise the demands placed on agricultural production, which can be  
13 interpreted as implying an increase in agricultural areas, an equivalent increase in productive  
14 efficiency (perhaps through greater inputs, i.e. higher intensity), or some combination of the two.  
15 Nonetheless, comparison with previous more complex model results suggests that the outcomes  
16 here are broadly equivalent. For example the vegan and vegetarian diets in Erb et al. (2016) have a  
17 central value for cropland area of approximately 1200 and 1000 Mha, respectively, compared to the  
18 low meat diet used here (based on the average diet in India) of 1022 Mha. As expected, for the  
19 reasons given above, changes in intensity considered in Erb et al. (2016) but not here appear to  
20 moderate the land use outcomes, i.e. for less agricultural land to be relinquished, but coupled with a  
21 decrease in intensity of production. Therefore, although the adopted approach neglects aspects that  
22 would allow robust spatial or temporal predictions of land use, it does provide a consistent  
23 methodology across scenarios allowing comparisons between them, a primary aim of the study.

24  
25 The results demonstrate that milk production is more efficient than the current animal product mix,  
26 with the milk scenario showing a decrease in land requirements (Table 1). Cull dairy cows and male  
27 dairy calves could also be used to produce beef, which is not accounted for in these results. If the  
28 additional beef production from an expanded dairy sector were considered, the land requirements in  
29 the milk scenario would be further reduced, as less land would be required to produce the remaining  
30 beef consumed. The magnitude of this bias is perhaps moderate, as the fraction of emissions from the  
31 dairy herd currently assigned to milk rather than meat production is between 90-96% (Opio et al.,  
32 2013).

### 33 34 *(b) Imitation meat and soybean production*

35 The imitation meat scenario, based on soybean curd, implies that more cropland is used for growing  
36 soybeans, while the other meat replacement scenarios use a more diverse mix of feeds. The  
37 additional soybean areas may be less suited to the crop and so would have lower yields than existing  
38 production, potentially leading to an underestimate of the area needed when using average yields.  
39 An additional 111 Mha of soybean area was calculated as needed (i.e. a doubling of 2013 area  
40 (FAOSTAT, 2015c)), while 248 Mha of cropland currently used for animal feed is spared. Therefore,  
41 the net cropland area decreases in this scenario suggest that suitable land may be available, although  
42 this would also be constrained by climatic suitability. However, higher soybean yields would be  
43 anticipated to have only a small impact on the results as the net percentage agricultural area change  
44 is dominated by the change in pasture area. The expansion of soybean area may have substantial  
45 local impacts, e.g. on biodiversity and soil quality, due to the intensity of production. However, the

1 land spared from agricultural production by the transition could be potentially used to offset such  
2 negative outcomes. This would be a form of 'land sparing', i.e. separation of land for conservation  
3 and food production, in contrast to 'land sharing' with integration of conservation and production  
4 (Phalan et al., 2011). However, attempting to account for the associated trade-offs and scale effects,  
5 as well as the challenges and controversy involved (Fischer et al., 2014), are out of scope for  
6 consideration here.

### 7 8 *(c) Cultured meat and energy*

9 The results suggest that the benefits claimed for cultured meat (Tuomisto and de Mattos, 2011) may  
10 not be justified. Although cultured meat was found to have a lower land footprint than beef, it had a  
11 similar efficiency to poultry meat (Figure 1 and 2), but with substantially higher direct energy  
12 requirements (Table 1 and S1). Direct energy inputs are needed for cultured meat to process raw  
13 biomass material into the cell medium, to then culture the cells and process them into a consumable  
14 product, including sterilisation and hydrolysis (Tuomisto and de Mattos, 2011). Conventional  
15 livestock use direct energy primarily in housing, e.g. lighting, heating and cooling (Macleod et al.,  
16 2013). Direct energy inputs for cultured meat (18-25 GJ/t (Tuomisto and de Mattos, 2011), Table 1)  
17 are higher than any of the other foods considered here (at least four times the highest conventional  
18 animal product, poultry meat (4.5 GJ/t (Macleod et al., 2013)). This suggests that a low-cost and low-  
19 carbon source of energy may be a prerequisite for cultured meat to be economically and  
20 environmentally viable. Furthermore, the provision of growth factors, vitamins and trace elements,  
21 e.g. B12, will also have an impact on the resources used for cultured meat, although the scale of this  
22 is unclear. However, the overall primary energy used in the production of cultured meat production  
23 was shown to be 46% lower than for beef production (e.g. including energy in fertiliser production  
24 and machinery), but 38% higher than for poultry meat. Given the relative novelty of this technology,  
25 further development and optimisation may be able to reduce these energy and cost requirements  
26 and increase the efficiency of production (Bhat et al., 2017). These improvements would potentially  
27 involve development of improved methods for producing the cell culture medium beyond that  
28 assumed here. The types of feed used may not match the current animal feed mix, although the land  
29 use consequences of such differences are likely to be lower than that associated with the uncertainty  
30 in efficiency of cultured meat production, and would not be expected to alter our conclusions.  
31 Overall, currently cultured meat could provide some benefits (e.g. land use savings compared to  
32 beef), but result in higher direct energy requirements and also potentially primary energy (e.g. in  
33 comparison to poultry meat). This conclusion concurs with a more recent anticipatory life cycle  
34 analysis of culture meat production (Mattick et al., 2015b).

### 35 36 *(d) Insects, promising but more research needed*

37 Insects are the most efficient animal production system considered, although less so than soybean  
38 curd. However, insects have the additional advantage that they are able to use a wide variety of  
39 feeds, including by-products and waste (Ocio and Vinaras, 1979; van Broekhoven et al., 2015). The  
40 results here assume that insect feed uses the same mix of feeds currently used for conventional  
41 livestock. However, if half of food discarded by consumers (from Alexander et al. (2017)) could be  
42 used as feed for mealworms, this would replace 8.1% of current animal production. Where the total  
43 feed is reduced there is potential for this to occur primarily for food commodities (e.g. cereals), and  
44 thereby increase the proportion of by-products. Although by-products are ascribed some value  
45 when considering their impacts (Elferink et al., 2008), the system efficiency increases by replacing

1 lower yielding conventional livestock with insects (Figure 1). For instance, soybeans could be used to  
2 produce soybean curd, and then feed insects from the residues.

3  
4 More research is needed to understand how the large scale production of insects could be achieved,  
5 the inputs required, the suitability of feeds, and other constraints (e.g. location) (van Huis, 2013).  
6 There is little published data on the feed efficiency of insect production. However direct energy  
7 inputs for intensive insect production appears comparable to intensive conventional livestock  
8 production (Oonincx and de Boer, 2012). Perhaps the biggest barrier to the large scale global  
9 adoption of insects as a food source is consumer acceptability (Looy et al., 2013; Shelomi, 2015),  
10 where again further research is required to understand how best to increase adoption and what rate  
11 and levels of consumption might be possible.

### 12 13 *(e) A future for ruminants?*

14 The land use footprint of ruminant meat production is high, and therefore consuming more beef and  
15 sheep meat requires large increases in land areas (Figure 2). Although ruminants are less efficient  
16 converters of feed to edible foods than monogastrics (Table 1), their high reliance on forage that is  
17 inedible to humans from non-arable land reduces their claim for feeds produced on cropland (Smil,  
18 2013). Livestock production can also provide a range of other benefits, e.g. recycling plant nutrients,  
19 maintaining ecosystems and providing social benefit (Janzen, 2011; Oltjen and Beckett, 1996).  
20 Therefore, ruminants that are mainly grass-fed from land that is unsuitable for the production of  
21 other crops may provide substantial benefits, but this implies a move away from intensive  
22 production practices, i.e. that use large quantities of feed produced from cropland. Such extensive  
23 grazing based systems are likely to produce a reduced quantity of livestock, and therefore per capita  
24 consumption rates of ruminant meat would have to continue to fall to avoid unsustainable land use  
25 change. Additionally, changes towards consumption of diets with lower land use requirements also  
26 provide the prospect of reduced competition for land between food production and climate change  
27 mitigation measures, e.g. bioenergy or afforestation (Smith et al., 2014).

## 28 29 **5. Conclusions**

30 These results suggest that alternatives to the current mix of livestock production systems could  
31 substitute current animal products and substantially reduce the current agricultural land use  
32 footprint from food production. Reducing meat consumption overall is likely to have the greatest  
33 effect on the land use footprint, but replacing beef or lamb with any of the foods considered here  
34 has the potential for substantial sustainability benefits. Although, the two most efficient products  
35 considered, i.e. imitation meat and insects, both come with consumer perception barriers, a shift  
36 towards poultry meat, eggs and milk was also found to offer land use and associated environmental  
37 benefits, of only slightly smaller magnitudes. Reductions in consumer waste have potentially  
38 important but smaller impacts on resource requirement than the other scenarios considered. We  
39 conclude that a diet which reduces agricultural land requirements may best be achieved through a  
40 combination of approaches, including both waste reduction, shifts towards more efficient  
41 conventional animal products (e.g. chicken and eggs), and greater use of alternatives such as insect  
42 and imitation meat. A more balanced approach than those in the stylised scenarios considered here  
43 would also require less extreme shifts in diets and therefore need less dramatic changes in consumer  
44 consumption habits. This work focuses principally on the land requirements, although out of scope

1 here, a similar consistent greenhouse gas lifecycle analysis across all options is warranted, as well as  
2 consideration of consequences for biodiversity, water requirements and other ecosystem services.  
3 Further research is also required into the technologies and production systems for the large scale  
4 production of insects, including what feeds are most appropriate and the potential use of food waste  
5 and by-products, and to better understand how consumer behaviour and preferences can be  
6 influenced towards a healthier and more sustainable diet.  
7

## 8 **6. References**

- 9 Alexander, P., Brown, C., Arneth, A., Finnigan, J., Moran, D., Rounsevell, M.D.A., 2017. Losses,  
10 inefficiencies and waste in the global food system. *Agricultural Systems* 153, 190–200.
- 11 Alexander, P., Brown, C., Rounsevell, M., Finnigan, J., Arneth, A., 2016. Human appropriation of land  
12 for food: The role of diet. *Global Environmental Change* 41, 88–98.
- 13 Alexander, P., Rounsevell, M.D.A., Dislich, C., Dodson, J.R., Engström, K., Moran, D., 2015. Drivers for  
14 global agricultural land use change: The nexus of diet, population, yield and bioenergy. *Global  
15 Environmental Change* 35, 138–147. doi:10.1016/j.gloenvcha.2015.08.011
- 16 Alexandratos, N., Bruinsma, J., 2012. *World Agriculture Towards 2030 / 2050*. FAO, Rome, Italy and  
17 IIASA, Laxenburg, Austria.
- 18 Bajželj, B., Richards, K.S., Allwood, J.M., Smith, P., Dennis, J.S., Curmi, E., Gilligan, C. a., 2014.  
19 Importance of food-demand management for climate mitigation. *Nature Climate Change* 4,  
20 924–929. doi:10.1038/nclimate2353
- 21 Bauer, C., Schlott, G., 2009. Fillet yield and fat content in common carp (*Cyprinus carpio*) produced in  
22 three Austrian carp farms with different culture methodologies. *Journal of Applied Ichthyology*  
23 25, 591–594. doi:10.1111/j.1439-0426.2009.01282.x
- 24 Bhat, Z.F., Bhat, H., Pathak, V., 2014. Prospects for In Vitro Cultured Meat – A Future Harvest, Fourth  
25 Edi. ed, *Principles of Tissue Engineering*. Elsevier. doi:10.1016/B978-0-12-398358-9.00079-3
- 26 Bhat, Z.F., Kumar, S., Bhat, H.F., 2017. In vitro meat: A future animal-free harvest. *Critical Reviews in  
27 Food Science and Nutrition* 57, 782–789. doi:10.1080/10408398.2014.924899
- 28 Bhat, Z.F., Kumar, S., Fayaz, H., 2015. In vitro meat production: Challenges and benefits over  
29 conventional meat production. *Journal of Integrative Agriculture* 14, 241–248.  
30 doi:10.1016/S2095-3119(14)60887-X
- 31 Bir, S., 2014. *From Poison to Passion: The Secret History of the Tomato, Modern Farmer*. Hudson, NY,  
32 USA.
- 33 Bodirsky, B.L., Rolinski, S., Biewald, A., Weindl, I., 2015. Global Food Demand Scenarios for the 21 st  
34 Century. *PLoS ONE*. doi:10.5281/zenodo.31008
- 35 Bostock, J., McAndrew, B., Richards, R., Jauncey, K., Telfer, T., Lorenzen, K., Little, D., Ross, L.,  
36 Handisyde, N., Gatward, I., Corner, R., 2010. *Aquaculture: global status and trends*.  
37 *Philosophical transactions of the Royal Society of London Series B, Biological sciences* 365,  
38 2897–2912. doi:10.1098/rstb.2010.0170
- 39 Bouvard, V., Loomis, D., Guyton, K.Z., Grosse, Y., Ghissassi, F. El, Benbrahim-Tallaa, L., Guha, N.,  
40 Mattock, H., Straif, K., 2015. Carcinogenicity of consumption of red and processed meat. *The  
41 Lancet Oncology* 16, 1599–1600. doi:10.1016/S1470-2045(15)00444-1
- 42 Bustamante, M., Robledo-Abad, C., Harper, R., Mbow, C., Ravindranath, N.H., Sperling, F., Haberl, H.,  
43 de Siqueira Pinto, A., Smith, P., 2014. Co-benefits, trade-offs, barriers and policies for  
44 greenhouse gas mitigation in the Agriculture, Forestry and Other Land Use (AFOLU) sector.  
45 *Global change biology* 44, 3270–3290. doi:10.1111/gcb.12591
- 46 De-Magistris, T., Pascucci, S., Mitsopoulos, D., 2015. Paying to see a bug on my food: How regulations  
47 and information can hamper radical innovations in the European Union. *British Food Journal*  
48 117, 1777–1792. doi:10.1108/BFJ-06-2014-0222
- 49 Dembosky, A., 2006. *How the Lobster Clawed its Way Up: A crustacean’s climb from pauper’s fare to  
50 modern-day delicacy*, Mother Jones. San Francisco, CA, USA.

1 Diana, J.S., 2009. Aquaculture Production and Biodiversity Conservation. *BioScience* 59, 27–38.  
2 doi:10.1525/bio.2009.59.1.7

3 Elferink, E. V., Nonhebel, S., Moll, H.C., 2008. Feeding livestock food residue and the consequences  
4 for the environmental impact of meat. *Journal of Cleaner Production* 16, 1227–1233.  
5 doi:10.1016/j.jclepro.2007.06.008

6 Engström, K., Olin, S., Rounsevell, M.D.A., Brogaard, S., van Vuuren, D.P., Alexander, P., Murray-Rust,  
7 D., Arneth, A., 2016a. Assessing uncertainties in global cropland futures using a conditional  
8 probabilistic modelling framework. *Earth System Dynamics Discussions* 7, 893–915.  
9 doi:10.5194/esd-2016-7

10 Engström, K., Rounsevell, M.D.A., Murray-Rust, D., Hardacre, C., Alexander, P., Cui, X., Palmer, P.I.,  
11 Arneth, A., 2016b. Applying Occam’s Razor to global agricultural land use change.  
12 *Environmental Modelling & Software* 75, 212–229. doi:10.1016/j.envsoft.2015.10.015

13 Erb, K.-H., Lauk, C., Kastner, T., Mayer, A., Theurl, M.C., Haberl, H., 2016. Exploring the biophysical  
14 option space for feeding the world without deforestation. *Nature communications*.  
15 doi:10.1038/pj.2016.37

16 FAO, 2016. *Fishery Statistical Collections: Global Aquaculture Production*. Food and Agriculture  
17 Organization of the United Nations (FAO), Rome, Italy.

18 FAO, 2012. *Livestock and Landscapes*. Food and Agriculture Organization of the United Nations  
19 (FAO), Rome, Italy.

20 FAO, 2006. *Livestock’s long shadow - environmental issues and options*. Food and Agriculture  
21 Organization of the United Nations (FAO), Rome, Italy. doi:10.1007/s10666-008-9149-3

22 FAOSTAT, 2015a. *Commodity Balances/Livestock and Fish Primary Equivalent (2015-12-16)*. Food and  
23 Agriculture Organization of the United Nations, Rome, Italy.

24 FAOSTAT, 2015b. *Production/Livestock Primary (2015-12-16)*. Food and Agriculture Organization of  
25 the United Nations, Rome, Italy.

26 FAOSTAT, 2015c. *Production/Crops (2015-12-16)*. Food and Agriculture Organization of the United  
27 Nations, Rome, Italy.

28 FAOSTAT, 2015d. *Commodity Balances/Crops Primary Equivalent (2015-12-16)*. Food and Agriculture  
29 Organization of the United Nations, Rome, Italy.

30 FAOSTAT, 2015e. *Food Supply - Livestock and Fish Primary Equivalent (2015-12-16)*. Food and  
31 Agriculture Organization of the United Nations, Rome, Italy.

32 FAOSTAT, 2015f. *Food Supply - Crops Primary Equivalent (2015-12-16)*. Food and Agriculture  
33 Organization of the United Nations, Rome, Italy.

34 FAOSTAT, 2015g. *Resources/Land (2015-12-16)*. Food and Agriculture Organization of the United  
35 Nations, Rome, Italy.

36 Finke, M.D., 2002. Complete nutrient composition of commercially raised invertebrates used as food  
37 for insectivores. *Zoo Biology* 21, 269–285. doi:10.1002/zoo.10031

38 Finnigan, T., Lemon, M., Allan, B., Paton, I., 2010. Mycoprotein, Life Cycle analysis and the food 2030  
39 Challenge. *Aspects of Applied Biology* 102.

40 Fischer, J., Abson, D.J., Butsic, V., Chappell, M.J., Ekroos, J., Hanspach, J., Kuemmerle, T., Smith, H.G.,  
41 von Wehrden, H., 2014. Land sparing versus land sharing: Moving forward. *Conservation Letters*  
42 7, 149–157. doi:10.1111/conl.12084

43 Gerber, P.J., Steinfeld, H., Henderson, B., Mottet, A., Opio, C., Dijkman, J., Falcucci, A., Tempio, G.,  
44 2013. Tackling climate change through livestock – A global assessment of emissions and  
45 mitigation opportunities. Food and Agriculture Organization of the United Nations (FAO), Rom.

46 Graça, J., Calheiros, M.M., Oliveira, A., 2015. Attached to meat? (Un)Willingness and intentions to  
47 adopt a more plant-based diet. *Appetite* 95, 113–125. doi:10.1016/j.appet.2015.06.024

48 Haberl, H., Erb, K.H., Krausmann, F., Bondeau, A., Lauk, C., Müller, C., Plutzer, C., Steinberger, J.K.,  
49 2011. Global bioenergy potentials from agricultural land in 2050: Sensitivity to climate change,  
50 diets and yields. *Biomass and Bioenergy* 35, 4753–4769. doi:10.1016/j.biombioe.2011.04.035

51 Havlík, P., Valin, H., Herrero, M., Obersteiner, M., Schmid, E., Rufino, M.C., 2014. Climate change  
52 mitigation through livestock system transitions. *Proceedings of the National Academy of*

1 Sciences 111, 3709–3714. doi:10.1073/pnas.1308044111  
2 Herrero, M., Conant, R., Havlik, P., Hristov, A.N., Smith, P., Gerber, P., Gill, M., Butterbach-Bahl, K.,  
3 Henderson, B., Valin, H., Thornton, P.K., 2016. Greenhouse gas mitigation potentials in the  
4 livestock sector. *Nature Climate Change* 6, 452–461. doi:10.1038/nclimate2925  
5 Hertel, T.W., Baldos, U.L.C., van der Mensbrugghe, D.Y., 2016. Predicting Long Term Food Demand,  
6 Cropland Use and Prices. *Annual Review of Resource Economics* 8.  
7 Hocquette, J., 2016. Is in vitro meat the solution for the future ? *MESC* 120, 167–176.  
8 doi:10.1016/j.meatsci.2016.04.036  
9 Houghton, R. a., House, J.I., Pongratz, J., van der Werf, G.R., DeFries, R.S., Hansen, M.C., Le Quéré, C.,  
10 Ramankutty, N., 2012. Carbon emissions from land use and land-cover change. *Biogeosciences*  
11 9, 5125–5142. doi:10.5194/bg-9-5125-2012  
12 Hu, F.B., 2011. Globalization of Diabetes: The role of diet, lifestyle, and genes. *Diabetes Care* 34,  
13 1249–1257. doi:10.2337/dc11-0442  
14 Humpenöder, F., Popp, A., Dietrich, J.P., Klein, D., Lotze-Campen, H., Bonsch, M., Bodirsky, B.,  
15 Weindl, I., Stevanovic, M., Müller, C., 2014. Investigating afforestation and bioenergy CCS as  
16 climate change mitigation strategies. *Environmental Research Letters* 9. doi:10.1088/1748-  
17 9326/9/6/064029  
18 Institute of Medicine, 2005. *Dietary Reference Intakes for Energy, Carbohydrate, Fiber, Fat, Fatty*  
19 *Acids, Cholesterol, Protein, and Amino Acids*. The National Academy Press, Washington, DC,  
20 USA. doi:10.1111/j.1753-4887.2004.tb00011.x  
21 Jamieson, S., 2015. Bug burgers and cricket crepes: Britain’s first insect restaurant opens in Wales.  
22 *The Telegraph*, 24 October 2015.  
23 Janzen, H.H., 2011. What place for livestock on a re-greening earth? *Animal Feed Science and*  
24 *Technology* 166–167, 783–796. doi:10.1016/j.anifeedsci.2011.04.055  
25 Kearney, J., 2010. Food consumption trends and drivers. *Philosophical transactions of the Royal*  
26 *Society of London Series B, Biological sciences* 365, 2793–807. doi:10.1098/rstb.2010.0149  
27 Keyzer, M. a., Merbis, M.D., Pavel, I.F.P.W., van Wesenbeeck, C.F. a., 2005. Diet shifts towards meat  
28 and the effects on cereal use: can we feed the animals in 2030? *Ecological Economics* 55, 187–  
29 202. doi:10.1016/j.ecolecon.2004.12.002  
30 Lamb, A., Green, R., Bateman, I., Broadmeadow, M., Bruce, T., Burney, J., Carey, P., Chadwick, D.,  
31 Crane, E., Field, R., Goulding, K., Griffiths, H., Hastings, A., Kasoar, T., Kindred, D., Phalan, B.,  
32 Pickett, J., Smith, P., Wall, E., zu Ermgassen, E.K.H.J., Balmford, A., 2016. The potential for land  
33 sparing to offset greenhouse gas emissions from agriculture. *Nature Climate Change*.  
34 doi:10.1038/nclimate2910  
35 Le Cotty, T., Dorin, B., 2012. A global foresight on food crop needs for livestock. *Animal : an*  
36 *international journal of animal bioscience* 6, 1528–36. doi:10.1017/S1751731112000377  
37 Leahy, E., Lyons, S., Tol, R.S.J., 2011. Determinants of vegetarianism and meat consumption  
38 frequency in Ireland. *Economic and Social Review* 42, 407–436.  
39 Le Quéré, C., Moriarty, R., Andrew, R.M., Peters, G.P., Ciais, P., Friedlingstein, P., Jones, S.D., 2015.  
40 *Global carbon budget 2014* 47–85. doi:10.5194/essd-7-47-2015  
41 Little, K., 2015. Burger chain adds bugs to the menu...on purpose. *CNBC*, 29 June 2015.  
42 Little, S., 2014. Feed Conversion Efficiency: A key measure of feeding system performance on your  
43 farm. *Dairy Australia*, Victoria, Australia.  
44 Looy, H., Dunkel, F. V., Wood, J.R., 2013. How then shall we eat? Insect-eating attitudes and  
45 sustainable foodways. *Agriculture and Human Values* 1–11. doi:10.1007/s10460-013-9450-x  
46 Macdiarmid, J.I., Douglas, F., Campbell, J., 2016. Eating like there’s no tomorrow: Public awareness of  
47 the environmental impact of food and reluctance to eat less meat as part of a sustainable diet.  
48 *Appetite* 96, 487–493. doi:10.1016/j.appet.2015.10.011  
49 Macleod, M., Gerber, P., Mottet, A., Tempio, G., Falcucci, A., Opio, C., Vellinga, T., Henderson, B.,  
50 Steinfeld, H., 2013. Greenhouse gas emissions from pig and chicken supply chains - A global life  
51 cycle assessment. *Food and Agriculture Organization of the United Nations (FAO)*, Rome, Italy.  
52 Malav, O.P., Talukder, S., Gokulakrishnan, P., Chand, S., 2015. Meat Analog: A Review. *Critical*

1           Reviews in Food Science and Nutrition 55, 1241–1245. doi:10.1080/10408398.2012.689381  
2 Mattick, C.S., Landis, A.E., Allenby, B.R., 2015a. A case for systemic environmental analysis of  
3           cultured meat. *Journal of Integrative Agriculture* 14, 249–254. doi:10.1016/S2095-  
4           3119(14)60885-6  
5 Mattick, C.S., Landis, A.E., Allenby, B.R., Genovese, N.J., 2015b. Anticipatory Life Cycle Analysis of In  
6           Vitro Biomass Cultivation for Cultured Meat Production in the United States. *Environmental*  
7           *Science and Technology* 49, 11941–9. doi:10.1021/acs.est.5b01614  
8 Meadu, V., Coche, I., Vermeulen, S., Friis, A.E., 2015. Paris Climate Agreement unlocks opportunities  
9           for food and farming. CGIAR, Copenhagen, Denmark.  
10 Mora, O., de Lattre-Gasquet, M., Donnars, C., Réchauchère, O., Le Mouël, C., Dumas, P., Barzman,  
11           M., Marty, P., Moreau, C., Brunelle, T., 2016. Scenarios of land use and food security in 2050,  
12           Agrimonde-Terra foresight. INRA, Paris, France.  
13 Moritz, M.S.M., Verbruggen, S.E.L., Post, M.J., 2015. Alternatives for large-scale production of  
14           cultured beef: A review. *Journal of Integrative Agriculture* 14, 208–216. doi:10.1016/S2095-  
15           3119(14)60889-3  
16 Mottet, A., Haan, C. De, Falcucci, A., Tempio, G., Opio, C., Gerber, P., 2017. Livestock : On our plates  
17           or eating at our table ? A new analysis of the feed / food debate. *Global Food Security* 1–8.  
18           doi:10.1016/j.gfs.2017.01.001  
19 Naylor, R.L., Hardy, R.W., Bureauc, D.P., Chiuva, A., Elliott, M., Farrell, A.P., Forster, I., Gatlin, D.M.,  
20           Goldburg, R.J., Hua, K., Nichols, P.D., 2009. Feeding aquaculture in an era of finite resources.  
21           *Proceedings of the National Academy of Sciences* 106, 15103–15110.  
22           doi:10.1073/pnas.0910577106  
23 NCD Risk Factor Collaboration, 2016. Trends in adult body-mass index in 200 countries from 1975 to  
24           2014: a pooled analysis of 1698 population-based measurement studies with 19.2 million  
25           participants. *Lancet* 387, 1377–1396. doi:10.1016/S0140-6736(16)30054-X  
26 Ocio, E., Vinaras, R., 1979. House fly larvae meal grown on municipal organic waste as a source of  
27           protein in poultry diets. *Animal Feed Science and Technology* 4, 227–231. doi:10.1016/0377-  
28           8401(79)90016-6  
29 Oltjen, J.W., Beckett, J.L., 1996. Role of Ruminant Livestock in Sustainable Agricultural Systems.  
30           *Journal of Animal Science* 74, 1406–1409.  
31 Oonincx, D.G.A.B., de Boer, I.J.M., 2012. Environmental Impact of the Production of Mealworms as a  
32           Protein Source for Humans - A Life Cycle Assessment. *PLoS ONE* 7, 1–5.  
33           doi:10.1371/journal.pone.0051145  
34 Opio, C., Gerber, P., Mottet, A., Falculli, A., Tempio, G., MacLeod, M., Vellinga, T., Henderson, B.,  
35           Steinfeld, H., 2013. Greenhouse gas emissions from ruminant supply chains- A global life cycle  
36           assessment. Food and Agriculture Organization of the United Nations (FAO), Rome, Italy.  
37 Pelletier, N., Tyedmers, P., 2010. Life cycle assessment of frozen tilapia fillets from Indonesian lake-  
38           based and pond-based intensive aquaculture systems. *Journal of Industrial Ecology* 14, 467–  
39           481. doi:10.1111/j.1530-9290.2010.00244.x  
40 Persijn, D., Charrondiere, U.R., 2014. Review of food composition data on edible insects. *FOOD*  
41           *CHEMISTRY*. doi:10.1016/j.foodchem.2014.10.114  
42 Phalan, B., Onial, M., Balmford, A., Green, R.E., 2011. Reconciling Food Production and Biodiversity  
43           Conservation: Land Sharing and Land Sparing Compared. *Science* 333, 1289–1291.  
44 Popkin, B.M., Carolina, N., Hill, C., 1999. Popkin(1999) Urbanization, Lifestyle Changes and the  
45           Nutrition. *World Development* 27, 1905–1916.  
46 Popkin, B.M., Gordon-Larsen, P., 2004. The nutrition transition: worldwide obesity dynamics and  
47           their determinants. *International journal of obesity and related metabolic disorders : journal of*  
48           *the International Association for the Study of Obesity* 28 Suppl 3, S2-9.  
49           doi:10.1038/sj.ijo.0802804  
50 Popp, A., Lotze-Campen, H., Bodirsky, B., 2010. Food consumption, diet shifts and associated non-  
51           CO2 greenhouse gases from agricultural production. *Global Environmental Change* 20, 451–462.  
52           doi:10.1016/j.gloenvcha.2010.02.001



- 1 Premalatha, M., Abbasi, T., Abbasi, T., Abbasi, S.A., 2011. Energy-efficient food production to reduce  
2 global warming and ecodegradation: The use of edible insects. *Renewable and Sustainable*  
3 *Energy Reviews* 15, 4357–4360. doi:10.1016/j.rser.2011.07.115
- 4 Ramos-Elorduy, J., 2009. Anthro-entomophagy: Cultures, evolution and sustainability.  
5 *Entomological Research* 39, 271–288. doi:10.1111/j.1748-5967.2009.00238.x
- 6 Rumpold, B. a., Schlüter, O.K., 2013. Potential and challenges of insects as an innovative source for  
7 food and feed production. *Innovative Food Science and Emerging Technologies* 17, 1–11.  
8 doi:10.1016/j.ifset.2012.11.005
- 9 SACN, 2011. Dietary Reference Values for Energy 2011. Scientific Advisory Committee on Nutrition,  
10 London, UK.
- 11 Sahirman, S., Ardiansyah, A., 2014. Assessment of tofu carbon footprint in banyumas, indonesia -  
12 towards “greener” tofu, in: *Proceeding of International Conference On Research,*  
13 *Implementation And Education Of Mathematics And Sciences 2014*, Yogyakarta State  
14 University, 18-20 May 2014.
- 15 Schaafsma, G., 2000. Criteria and Significance of Dietary Protein Sources in Humans The Protein  
16 Digestibility – Corrected Amino Acid Score 1 1865–1867.
- 17 Schmitz, C., van Meijl, H., Kyle, P., Nelson, G.C., Fujimori, S., Gurgel, A., Havlik, P., Heyhoe, E., d’Croz,  
18 D.M., Popp, A., Sands, R., Tabeau, A., van der Mensbrugge, D., von Lampe, M., Wise, M., Blanc,  
19 E., Hasegawa, T., Kavallari, A., Valin, H., 2014. Land-use change trajectories up to 2050: insights  
20 from a global agro-economic model comparison. *Agricultural Economics* 45, 69–84.  
21 doi:10.1111/agec.12090
- 22 Shelomi, M., 2015. Why we still don’t eat insects: Assessing entomophagy promotion through a  
23 diffusion of innovations framework. *Trends in Food Science & Technology* 45, 1–8.  
24 doi:10.1016/j.tifs.2015.06.008
- 25 Smil, V., 2013. *Should We Eat Meat? Evolution and Consequences of Modern Carnivory*. Wiley, New  
26 York, USA.
- 27 Smith, K.A., 2013. Why the Tomato Was Feared in Europe for More Than 200 Years: How the fruit got  
28 a bad rap from the beginning. *Smithsonian*.
- 29 Smith, P., 2013. Delivering food security without increasing pressure on land. *Global Food Security* 2,  
30 18–23. doi:10.1016/j.gfs.2012.11.008
- 31 Smith, P., Bustamante, M., Ahammad, H., Clark, H., Dong, H., Elsidig, E., Tubiello, F., Smith P.M.  
32 Bustamante, H. Ahammad, H. Clark, H. Dong, E.A. Elsidig, H. Haberl, R. Harper, J. House, M.  
33 Jafari, O.M., C. Mbow, N.H. Ravindranath, C.W. Rice, C. Robledo Abad, A. Romanovskaya, F.  
34 Sperling, and F.T., 2014. Agriculture, Forestry and Other Land Use (AFOLU), in: [Edenhofer, O.,  
35 R., Pichs-Madruga, Y. Sokona, E. Farahani, S. Kadner, K. Seyboth, A. Adler, I. Baum, S. Brunner,  
36 P. Eickemeier, B. Kriemann, J., Savolainen, S. Schlömer, C. von Stechow, T.Z. and J.C.M. (Eds.),  
37 *Climate Change 2014: Mitigation of Climate Change. Contribution of Working Group III to the*  
38 *Fifth Assessment Report of the Intergovernmental Panel on Climate Change*. Cambridge  
39 University Press, Cambridge, UK and NY, USA, pp. 811–922.
- 40 Smith, P., Gregory, P.J., 2013. Climate change and sustainable food production. *The Proceedings of*  
41 *the Nutrition Society* 72, 21–8. doi:10.1017/S0029665112002832
- 42 Spang, B., 2013. Insects as food: Assessing the food conversion efficiency of the mealworm (*Tenebrio*  
43 *molitor*). Evergreen State College, WA, USA.
- 44 Stehfest, E., Bouwman, L., Van Vuuren, D.P., Den Elzen, M.G.J., Eickhout, B., Kabat, P., 2009. Climate  
45 benefits of changing diet. *Climatic Change* 95, 83–102. doi:10.1007/s10584-008-9534-6
- 46 Tabassum-Abbasi, Abbasi, T., Abbasi, S.A., 2016. Reducing the global environmental impact of  
47 livestock production : the minilivestock option. *Journal of Cleaner Production* 112, 1754–1766.  
48 doi:10.1016/j.jclepro.2015.02.094
- 49 Tacon, A.G.J., Metian, M., 2008. Global overview on the use of fish meal and fish oil in industrially  
50 compounded aquafeeds: Trends and future prospects. *Aquaculture* 285, 146–158.  
51 doi:10.1016/j.aquaculture.2008.08.015
- 52 Thornton, P.K., 2010. Livestock production: recent trends, future prospects. *Philosophical*

1 transactions of the Royal Society of London Series B, Biological sciences 365, 2853–2867.  
2 doi:10.1098/rstb.2010.0134

3 Tilman, D., Balzer, C., Hill, J., Befort, B.L., 2011. Global food demand and the sustainable  
4 intensification of agriculture. *Proceedings of the National Academy of Sciences of the United*  
5 *States of America* 108, 20260–4. doi:10.1073/pnas.1116437108

6 Townsend, E., 2012. *Lobster: A global history*. University of Chicago Press, Chicago, IL, USA.

7 Tuomisto, H.L., de Mattos, M.J.T., 2011. Environmental impacts of cultured meat production.  
8 *Environmental science & technology* 45, 6117–23. doi:10.1021/es200130u

9 UNFCCC, 2015. COP21: Adoption of the Paris Agreement. United Nations Framework Convention on  
10 Climate Change.

11 USDA, 2015. National Nutrient Database for Standard Reference Release 28. United States  
12 Department of Agriculture, Agricultural Research Service.

13 van Broekhoven, S., Oonincx, D.G. a. B., van Huis, A., van Loon, J.J. a., 2015. Growth performance and  
14 feed conversion efficiency of three edible mealworm species (Coleoptera: Tenebrionidae) on  
15 diets composed of organic by-products. *Journal of Insect Physiology* 73, 1–10.  
16 doi:10.1016/j.jinsphys.2014.12.005

17 van Huis, A., 2013. Potential of Insects as Food and Feed in Assuring Food Security. *Annual Review of*  
18 *Entomology* 58, 563–83. doi:10.1146/annurev-ento-120811-153704

19 Verbeke, W., Marcu, A., Rutsaert, P., Gaspar, R., Seibt, B., Fletcher, D., Barnett, J., 2015. “Would you  
20 eat cultured meat?”: Consumers’ reactions and attitude formation in Belgium, Portugal and  
21 the United Kingdom. *Meat Science* 102, 49–58. doi:10.1016/j.meatsci.2014.11.013

22 Verstrate, P., 2016. Feeding the 7 Billion: Cultured meat, in: *Edinburgh International Science Festival,*  
23 *Summerhall, Edinburgh, UK.*

24 Vinnari, M., Mustonen, P., Räsänen, P., 2010. Tracking down trends in non-meat consumption in  
25 Finnish households, 1966-2006. *British Food Journal* 112, 836–852.  
26 doi:10.1108/00070701011067451

27 Wang, H.L., Cavins, J.F., 1989. Yield and amino acid composition of fractions obtained during tofu  
28 production. *Cereal chemistry (USA)* 66, 359–361.

29